

1 Latent Gaussian models to predict historical bycatch in  
2 commercial fishery

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14 Knowledge about how many fish that have been killed due to bycatch is an important  
15 aspect of ensuring a sustainable ecosystem and fishery. We introduce a Bayesian spatio-  
16 temporal prediction method for historical bycatch that incorporates two sources of available  
17 data sets, fishery data and survey data. The model used assumes that occurrence of bycatch  
18 can be described as a log-linear combination of covariates and random effects modeled as  
19 Gaussian fields. Integrated Nested Laplace Approximations (INLA) is used for fast calcula-  
20 tions. The method introduced is general, and is applied on bycatch of juvenile cod (*Gadus*  
21 *morhua*) in the Barents Sea shrimp (*Pandalus borealis*) fishery. In this fishery we compare  
22 our prediction method with the well known ratio and effort methods, and make a strong  
23 case that the Bayesian spatio-temporal method produces more reliable historical bycatch  
24 predictions compared to existing methods.

25 **Keywords:** Bycatch, Spatio-temporal, Bayesian, INLA, Commercial fishery

## 26 1 Introduction

27 Bycatch in commercial fisheries may potentially threaten a sustainable ecosystem and fishery,  
28 and knowledge about historical bycatch is therefore important. If bycatch is not recorded in the  
29 fishermen catch logbooks, which is the main source of information within commercial fisheries,  
30 historical bycatch needs to be estimated. In this research, we introduce a prediction procedure  
31 based on the newly constructed Bayesian hierarchical spatio-temporal bycatch model in Breivik  
32 et al. (2016). We further compare our method with the frequently used ratio method (Scheaffer  
33 et al., 1996, page 204) and effort method (e.g. Walmsley et al., 2007; Hall, 1996) for a specific  
34 fishery.

35 Typically two sources of data are available for predicting bycatch; the commercial catch logbooks  
36 the fishermen are obliged to report, and observations taken for monitoring purposes. The first  
37 source, referred to as fishery data, contains only target catch, while the latter, referred to  
38 as survey data, contains both target catch and bycatch. To predict historical bycatch in the  
39 commercial fishery, we combine the fishery data with the survey data.

40 The ratio method and the effort based method are widely used to predict historical bycatch

41 (Davies et al., 2009; Vinther, 1999; Ye et al., 2000; Amandè et al., 2010; Ye, 2002; Walmsley  
42 et al., 2007). The ratio method scales the commercial target catch with the observed bycatch  
43 ratio in the survey data, while the effort based method scales the observed bycatch with the  
44 commercial trawl effort.

45 The model proposed to predict historical bycatch takes a regression approach and utilizes possi-  
46 ble important explanatory variables (such as seasonal effects and the size of target catch). It also  
47 includes an underlying stochastic structure that partly explains the processes that the explana-  
48 tory variables fail to capture and simultaneously takes dependence structures into account. By  
49 using our bycatch model we can utilize observations taken over several years to describe global  
50 structures of bycatch. Our model-based approach is thereby able to provide good realistic by-  
51 catch predictions (with uncertainty) even in areas and time periods with few or no inspected  
52 trawl hauls.

53 The prediction method introduced in this research is general and is applied to bycatch of juvenile  
54 cod in the Barents Sea shrimp fishery. A sorting grid, which sorts out the larger cod and reduces  
55 bycatch, was imposed in this fishery in 1992/1993 (ICES, 1994). Because of the grid, the  
56 bycatch is of no commercial value, and is discarded. There is a real time regulation of this  
57 fishery with respect to bycatch of juvenile cod, haddock (*Melanogrammus aeglefinus*), redfish  
58 (*Sebastes norvegicus* and *Sebastes mentella*), Greenland halibut (*Reinhardtius hippoglossoides*)  
59 and undersized shrimp. If the Norwegian Directorate of Fisheries Monitoring and Surveillance  
60 Service (MSS) believes that an area has a higher bycatch ratio than allowed, that is e.g. 8  
61 cod per 10 kilogram of shrimps (Fiskeridirektoratet, 2005), the area is temporarily closed. The  
62 survey data used in this research have previously been used by MSS to regulate the shrimp  
63 fishery (Breivik et al., 2016). See Little et al. (2015) for a summary of management methods  
64 with respect to bycatch in several other large fisheries.

65 Bycatch was also predicted in Breivik et al. (2016) for regulation purposes. Our research differs  
66 mainly because we utilize huge amounts of fishery data, resulting in new computational diffi-  
67 culties, and that the data distribution is changed from log-Gaussian to zero-inflated negative  
68 binomial. Furthermore, the target catch is in this research a given covariate since it is included  
69 in both the fishery data and the survey data, while in Breivik et al. (2016) where future predic-  
70 tions was the focus, the shrimp catch was stochastic. To adapt to the information given in the

71 fishery data, the response variable for bycatch in Breivik et al. (2016) is changed from *bycatch*  
 72 *per nautical mile* to *total bycatch*, and with duration trawled included as an offset.

73 The paper is organized as follows. Section 2 presents the data used for historical bycatch  
 74 prediction. Section 3 provides a brief overview of historical bycatch prediction methods. Section  
 75 4 presents the model and section 5 illustrates the inference and prediction procedure. Section  
 76 6 presents the estimated model and predictions of historical bycatch. Section 7 validates the  
 77 predictions and compares them with the ratio and effort method. Finally, section 8 and 9 present  
 78 discussion and conclusions.

## 79 2 Data

80 Figure 1 shows the spatial distribution of the data. The left panel shows the spatial resolution  
 81 of the fishery data (specific locations are not recorded), while the right panel shows the spatial  
 82 locations of the survey data.

83 There were reported in total 81,809 commercial shrimp catches during the period 1994 to 2006.  
 84 Table 1 gives a short summary of possible covariates in the fishery data. Notice that the fishery  
 85 data consists of daily catches, meaning that if a vessel has made several trawl hauls in the same  
 86 small-scale spatial unit (see Figure 1) in a single day, this counts as one record.

Data	Description
Time	Date of catch (day, month and year)
Location	Which region the catch was taken (see small areas in Figure 1a)
Target catch	Total shrimp catch by one boat in a given area and day (770kg, 13,750kg)
Duration	Hours used to trawl by a boat in a given area and day (7 hours, 22.9 hours)
Number of trawls	The number of trawls varies between (76%), two (23%) or three (1.7%)
Quarter of the year	1st (9.2%), 2nd (42%), 3rd (38%) and 4th (11%)

Table 1: *Summary of fishery data, intervals in parentheses are 90% coverage intervals.*

87 We used 7,363 observations of shrimp and bycatch of cod from 1994 to 2006 taken by the MSS  
 88 (the survey data), and provided by the Institute of Marine Research (IMR) in Bergen, Norway,  
 89 see Table 2 for a short summary of the survey data. There were 18.5% zero-observations of  
 90 bycatch. The survey observations are collected for regulation purposes and the trawl hauls are  
 91 conducted using the same equipment as in the commercial fishery. These observations may

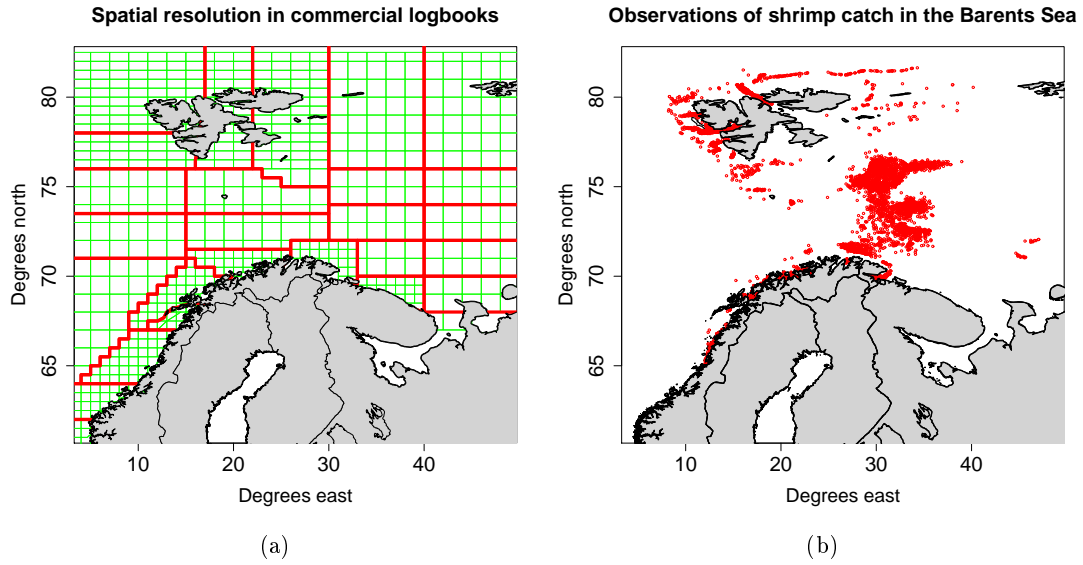


Figure 1: a) Map of the Barents Sea with small green rectangles describing the spatial resolution of the fishery data. The larger red areas are used when calculating the ratio and effort estimates. b) Map of the Barents Sea with red dots illustrating the survey data.

Data	Description
Target catch	Shrimp catch varied between 2.4 kilogram and 17.7 tons (20, 3,190)
Bycatch cod	The number of cod varied between 0 and 35,775 cod (0, 1,008)
Time	Time of catch down to minutes scale
Location	Catch location (single point) given in longitude and latitude
Open/Closed	Describes if the location was open for commercial fishery or not (83% open)
Duration trawled	The hours used to trawl (1.6 hours, 6 hours)
Number of trawls	The number of trawls varies between one (74%), two (23%) or three (3.0%)
Temperature	Bottom sea temperature (0.17, 9.3)
Depth	Ocean depth at catch location (227, 410)
Quarter of the year	1st (21%), 2nd (35%), 3rd (20%) and 4th (23%)

Table 2: Summary of data collected by the MSS, intervals in parentheses are 90% coverage intervals.

92 either have been taken on board vessels active in the commercial fishery (23%), or by vessels  
 93 hired by the MSS (77%) for collecting a sufficient amount of observations at selected areas where  
 94 commercial shrimp trawling occurs.

95 In addition to the variables in Table 1 we also use total abundance estimates of 0-group cod  
 96 (juvenile cod less than one year old) in the whole Barents Sea to predict the historical bycatch.  
 97 These estimates can be found in Jakobsen and Ozhigin (2011, pages 565-567).

### 98 **3 Methods to estimate historical bycatch**

99 This section gives a brief overview of methods to estimate historical bycatch. Our research  
100 focuses on the third method (the model based method).

#### 101 **3.1 The ratio method**

102 The ratio method (Scheaffer et al., 1996, page 204) has been widely used to estimate historical  
103 bycatch. The ratio method uses the reported bycatch ratio in the survey data to scale the  
104 commercial target catch (here shrimp) to achieve estimates of bycatch, and is defined as

$$\widehat{B}_{A,t}^{\text{ratio}} = \frac{\sum_{i=1}^n b_{i,A,t}}{\sum_{i=1}^n z_{i,A,t}} Z_{A,t} = R_{A,t} Z_{A,t}. \quad (1)$$

105 Here  $(z_{i,A,t}, b_{i,A,t})$  are the  $i$ th observed target catch and bycatch in the survey data in area  $A$  and  
106 time interval  $t$ ,  $Z_{A,t}$  is the total commercial target catch in area  $A$  and time interval  $t$ , and  $R_{A,t}$   
107 is the observed bycatch ratio in area  $A$  and time interval  $t$ . The historical bycatch in several  
108 time intervals can then be estimated in the whole Barents Sea as  $\sum_A \sum_t R_{A,t} Z_{A,t}$ . We let the  
109 areas,  $A$ , be the small green rectangles in Figure 1a and each time intervals,  $t$ , be quarters of  
110 years. The ratio method with these areas and time intervals is currently used as a standard for  
111 providing official historical bycatch estimates in the Barents Sea shrimp fishery (Ajiad et al.,  
112 2007; Hysten and Jacobsen, 1987).

113 Equation (1) assumes there exists survey data in each area and time interval where commercial  
114 catches occurred. This is not always fulfilled, and in such situations it is a common procedure  
115 to expand the area on which the ratio,  $R_{A,t}$ , is calculated. In our experiments, we expand the  
116 area in the following order: First we use all observations in the larger red area containing the  
117 area of interest (Figure 1a) within the given time interval. If there are no observations in this  
118 larger area, we use all the observations in the Barents Sea within the given time interval. If  
119 there are no observations in the Barents Sea, we use all observations collected one time interval  
120 before and after. We also experimented with expanding the time interval before increasing the  
121 spatial areas, but this had little effect on the results. Our first expansion step is similar to the

122 one used in Ajiad et al. (2007), but the next expansion steps were not documented in detail  
 123 in Ajiad et al. (2007). Furthermore, as done in Ajiad et al. (2007), only observations taken at  
 124 locations open for commercial fishery is used to calculate the bycatch ratio (1).

### 125 **3.2 The effort method**

126 Another much used method for estimating historical bycatch is the effort method (e.g. Walmsley  
 127 et al., 2007). The effort method uses reported trawl effort in the commercial fishery to up-scale  
 128 bycatch estimates from the survey data, and is defined as

$$\hat{B}_{A,t}^{\text{effort}} = \frac{\sum_{i=1}^n b_{i,A,t}}{\sum_{i=1}^n \text{time}_{i,A,t}} T_{A,t}. \quad (2)$$

129 Here  $\text{time}_{i,A,t}$  is towing time used when  $b_{i,A,t}$  was observed, and  $T_{A,t}$  is the total commercial  
 130 trawl time within area  $A$  and time interval  $t$ . Note that this method is (at this time) not used  
 131 for estimating historical bycatch in the Barents Sea shrimp fishery (Ajiad et al., 2007; Hylen  
 132 and Jacobsen, 1987), but we include it in this research since it is a natural alternative to the  
 133 ratio method in this fishery.

134 The effort method (2) also assumes there exists survey data in each area and time interval where  
 135 commercial catches occurred. When this is not fulfilled, we increase the area, and potentially  
 136 time, as described for the ratio method. Just as for the ratio method (1), only observations  
 137 taken at locations open for commercial fishery is used to calculate the effort estimate (2).

### 138 **3.3 A model-based procedure**

139 A model-based procedure constructs a model for the observed bycatch and uses the model to  
 140 estimate the unobserved historical bycatch. Let  $\mathbf{B}_C$  and  $\mathbf{B}_S$  be the bycatch from the fishery  
 141 data and the survey data, respectively. We know  $\mathbf{B}_S$  and want to estimate  $\mathbf{B}_C$ . Let further  
 142  $\mathbf{Z} = (\mathbf{Z}_C, \mathbf{Z}_S)$  be the target catch from both fishery data and the survey data. By using a  
 143 probabilistic model,  $M$ , we can focus on the distribution

$$P(\mathbf{B}_C | \mathbf{B}_S, \mathbf{Z}, M), \quad (3)$$

144 and use this distribution to construct predictions of historical bycatch with uncertainty.

145 As opposed to the two previous methods, the model based method (3) does not assume there  
 146 exist survey data in each area and time interval where commercial catches occur. However,  
 147 for the model to give realistic predictions, it is crucial that it is able to utilize other sources  
 148 of information such as relevant explanatory variables and dependence structures. Unlike the  
 149 ratio (1) and effort method (2), the model-based procedure (3) is able to utilize survey data at  
 150 locations closed for commercial fishery in order to predict historical bycatch.

## 151 4 The model

152 In this section we introduce our model for historical bycatch (3). The model is a modified version  
 153 of that introduced in Breivik et al. (2016). Let  $B(\mathbf{s}, t)$  be the number of juvenile cod caught at  
 154 time  $t$  and location  $\mathbf{s}$ . We model  $B(\mathbf{s}, t)$  as zero-inflated negative binomial distributed, that is  
 155 with density

$$\pi(B(\mathbf{s}, t)) = p(\mu(\mathbf{s}, t))\mathbb{I}_{B(\mathbf{s}, t)=0} + [1 - p(\mu(\mathbf{s}, t))]\text{NB}(B(\mathbf{s}, t); \mu(\mathbf{s}, t), \varsigma). \quad (4)$$

156 Here  $p(\mu)$  represent an additional probability for zero,  $\mathbb{I}_D$  is an indicator function which is equal  
 157 to one if  $D$  is true and zero otherwise, and  $\text{NB}(\cdot; \mu, \varsigma)$  is the negative binomial density with  
 158 expectation  $\exp(\mu)$  and dispersion parameter  $\varsigma$ . The log-expectation,  $\mu(\mathbf{s}, t)$ , of the negative  
 159 binomial distribution in (4) is modeled as:

$$\mu(\mathbf{s}, t) = \mathbf{X}(\mathbf{s}, t)^T \boldsymbol{\beta} + \alpha(\mathbf{s}) + v(t) + \gamma(\mathbf{s}, t), \quad (5)$$



160 where  $\mathbf{X}(\mathbf{s},t)$  is a vector of covariates and  $\boldsymbol{\beta}$  the vector of corresponding regression coefficients.  
 161 Three random effect terms are included in the expectation, one spatial,  $\alpha(\mathbf{s})$ , one temporal,  $v(t)$ ,  
 162 and one spatio-temporal,  $\gamma(\mathbf{s},t)$ . These are respectively intended to capture that the bycatch  
 163 amounts may depend on local features, that bycatch changes between years and that observations  
 164 close to each other in both space and time are highly correlated. The random effects are modeled  
 165 as Gaussian random fields.

166 The additional zero-probability,  $p(\mu)$ , in (4) is modeled as

$$p(\mu(\mathbf{s},t)) = 1 - \left( \frac{\exp(\mu(\mathbf{s},t))}{1 + \exp(\mu(\mathbf{s},t))} \right)^a, \quad (6)$$

167 where  $a > 0$  and adjusts how the zero-probability changes with respect to (5).

#### 168 4.1 Covariates

169 The covariates that have been considered are given in Table 3. Notice that shrimp catch is  
 170 in this setting a given covariate, and differs from the model in Breivik et al. (2016) where the  
 171 shrimp catch was considered stochastic. In Breivik et al. (2016) the time of the day was also  
 172 found important for predicting bycatch, but this variable is not given in the fishery data and is  
 173 therefore not used in this research. We use estimated abundance of 0-group cod in the whole  
 174 Barents Sea as a covariate. Breivik et al. (2016) tried to utilize the spatial locations of the 0-  
 175 group estimates as a spatial predictor, but did not find support in the data for such a procedure.  
 176 Note that the number of trawls used at the same time is included as a categorical variable and  
 177 not as an offset, this is done since the shape of the trawl may vary with the number of trawls  
 178 used at the same time.

179 We use a Fourier series (Lay, 2006, page 456) for the seasonal effect. The Fourier series is given  
 180 by

$$f(t') = \sum_{i=1}^r (c_i \sin(it') + d_i \cos(it')), \quad (7)$$

Covariates	Type	Description
0-group	Continuous	Logarithm of aggregated 0-group abundance of cod
Temperature (standardized)	Continuous	Bottom sea temperature
Depth (standardized)	Continuous	Ocean depth at catch location
Target catch	Continuous	Logarithm of hourly shrimp catch
Number of trawls	Categorical	The number of trawls used at the same time
Seasonal effect	Continuous	Fourier series (7)
Time (scaled to years)	Continuous	Linear covariate of time
Duration	Continuous	Duration of trawl (used as offset)

Table 3: *Covariates considered.*

181 were  $t' \in [0, 2\pi]$  is a linear function of time such that  $t' = 0$  for 1st January and  $t' = 2\pi$  for 31st  
182 December. The parameters  $c_i$  and  $d_i$  in (7) correspond to regression coefficients in (5), and  $r$  is  
183 the number of harmonics in the Fourier series. As in Breivik et al. (2016), we allow the seasonal  
184 effect to be a function of latitude to accommodate for different cod growth ratios which depends  
185 on temperature (see Breivik et al. (2016) for details).

## 186 4.2 Correlation structure

187 We assume as in Breivik et al. (2016) that the spatially correlated Gaussian field in (5),  $\alpha(\mathbf{s})$ ,  
188 follows a stationary Matern covariance structure:

$$\text{Cov}(\alpha(\mathbf{s}_1), \alpha(\mathbf{s}_2)) = \frac{\sigma_\alpha^2}{2^{\nu-1}\Gamma(\nu)} (\kappa_\alpha \|\mathbf{s}_1 - \mathbf{s}_2\|)^\nu K_\nu(\kappa_\alpha \|\mathbf{s}_1 - \mathbf{s}_2\|), \quad (8)$$

189 where  $\sigma_\alpha^2$  is the marginal variance,  $\|\cdot\|$  is the Euclidean distance measure in kilometers,  $\nu$  is a  
190 smoothing parameter,  $\kappa_\alpha$  is a spatial scale parameter and  $K_\nu(\cdot)$  is the modified Bessel function  
191 of the second kind. As in Breivik et al. (2016) we fix  $\nu = 1$  since this value is typically poorly  
192 identifiable (Blangiardo and Cameletti, 2015, page 194).

193 We assume as in Breivik et al. (2016) the time-dependent zero-mean Gaussian random field,  $v(t)$ ,  
194 to be constant within years while independent between years, with variance  $\sigma_v^2$ . We further define  
195 the first month of the year to be September when we refer to the yearly effect. This is reasonable  
196 because the 0-group enters a demersal life stage after September, and thereby starts living on the  
197 seabed where shrimp trawling occurs (Jakobsen and Ozhigin, 2011, page 230). Note that this

198 temporal structure comes in addition to possible linear time trend and seasonal effects.

199 The spatio-temporal interaction term,  $\gamma(\mathbf{s}, t)$ , is modeled with mean zero and a separable sta-  
200 tionary exponential covariance structure given by

$$\text{Cov}\left(\gamma(\mathbf{s}_1, t_1), \gamma(\mathbf{s}_2, t_2)\right) = \sigma_\gamma^2 \exp\left(-\frac{\|\mathbf{s}_1 - \mathbf{s}_2\|}{\theta_s} - \frac{|t_1 - t_2|}{\theta_t}\right). \quad (9)$$

201 Here  $\sigma_\gamma^2$  is the marginal variance,  $\|\cdot\|$  is the Euclidean distance measure in kilometers,  $|t_1 - t_2|$   
202 is the time difference in days and  $\theta_s$  and  $\theta_t$  are range parameters in space and time.

## 203 5 Inference and prediction procedure

204 This section elaborates the inference and prediction procedure, and is divided into two subsec-  
205 tion. The first subsection elaborates the inference, while the second subsection elaborates the  
206 prediction procedure. Note that only survey data are used for inference, and the fishery data  
207 are used combined with the survey data for prediction.

### 208 5.1 Inference

209 Only the survey data are used for inference on models and model parameters. The Bayesian  
210 inference is performed with the integrated nested Laplace approximation (INLA) technique (Rue  
211 et al., 2009; Martins et al., 2013) with use of the R-package R-INLA (<http://www.r-inla.org>).  
212 The INLA technique is an efficient procedure for fast approximation of the parameters and latent  
213 fields in the model. Non-informative priors are used, see appendix A.1, and we refer to Breivik  
214 et al. (2016) for more details on the inference procedure.

215 Which correlation structures to include is first selected with use of all the relevant covariates.  
216 The covariates are then selected with a backwards elimination procedure given the selected  
217 correlation structure. This ordering for selecting parameters is the same as in Breivik et al.  
218 (2016); Zuur (2009, page 121).

219 We have used the Bayes factor (Gelfand, 1996) for selection of correlation structures and covari-  
220 ates. In Breivik et al. (2016) three other validation criteria were used to evaluate the covariance

221 structure in the model for bycatch of cod. Then all the model selection criteria agreed and  
 222 we believe it is satisfactory to only use the Bayes factor in this research. The Bayes factor is  
 223 the ratio of the marginal likelihoods (ML) given by  $ML = P(\mathbf{B}_S|M)$ . See Rue et al. (2009) on  
 224 how the ML is calculated within R-INLA. Our model selection procedure has one exception.  
 225 The 0-group regression parameter is highly confounded with the yearly effect by construction.  
 226 Because of this the marginal likelihood is not adequate for selection of the 0-group when the  
 227 yearly effect is included. Just as in Breivik et al. (2016), if the yearly effect is included, the  
 228 0-group is included if it has predictive power.

## 229 5.2 Historical bycatch prediction

230 The historical bycatch is predicted by first fitting the selected model from section 5.1 with the  
 231 survey data using R-INLA, and then, based on the given estimated model, using a prediction  
 232 procedure which samples from the posterior distribution. This subsection elaborates on the  
 233 historical bycatch prediction.

234 Let  $\boldsymbol{\varphi} = \{\varphi(\mathbf{s},t)\}$  be the vector of latent fields where

$$\varphi(\mathbf{s},t) = \alpha(\mathbf{s}) + \nu(t) + \gamma(\mathbf{s},t) \quad (10)$$

235 if all fields are included in the model (5), while some of the terms can be missing in general. Let  
 236 also  $\boldsymbol{\varphi}_C$  and  $\boldsymbol{\varphi}_S$  be the subvectors of  $\boldsymbol{\varphi}$  corresponding to the commercial bycatch and surveillance  
 237 bycatch. The latent structure is of the form

$$\begin{pmatrix} \boldsymbol{\varphi}_C \\ \boldsymbol{\varphi}_S \end{pmatrix} \sim N(\mathbf{0}, \boldsymbol{\Sigma}) = N \left( \begin{pmatrix} \mathbf{0} \\ \mathbf{0} \end{pmatrix}, \begin{pmatrix} \boldsymbol{\Sigma}_{CC} & \boldsymbol{\Sigma}_{CS} \\ \boldsymbol{\Sigma}_{SC} & \boldsymbol{\Sigma}_{SS} \end{pmatrix} \right), \quad (11)$$

238 where  $\boldsymbol{\Sigma}$  represents the selected covariance structure with sub-elements  $\boldsymbol{\Sigma}_{CC}, \boldsymbol{\Sigma}_{CS}, \boldsymbol{\Sigma}_{SC}$  and  
 239  $\boldsymbol{\Sigma}_{SS}$  defining respectively the correlation between the commercial bycatch, the cross correla-  
 240 tion between the commercial bycatch and the surveillance bycatch and the correlation between  
 241 the surveillance bycatch. All these terms are derived from the set of latent fields that are

242 included in the model. Note that we do not know the exact locations of the fishery data,  
 243  $\mathbf{L} = \{(\mathbf{s}, t) : (\mathbf{s}, t) \text{ corresponds to fishery data locations}\}$ , needed in the covariance structure. To  
 244 account for the uncertainty in  $\mathbf{L}$ , we assume for simplicity that the fishery data are independent  
 245 and uniformly distributed on the areas reported (the green rectangles in Figure 1a).

246 The distribution of the commercial bycatch given the survey data is given by

$$\pi(\mathbf{B}_C | \mathbf{B}_S) = \int \pi(\mathbf{B}_C | \boldsymbol{\beta}, \boldsymbol{\varphi}_C, \boldsymbol{\theta}) \pi(\boldsymbol{\varphi}_C | \boldsymbol{\theta}, \boldsymbol{\varphi}_S, \mathbf{L}) \pi(\boldsymbol{\theta}, \boldsymbol{\beta}, \boldsymbol{\varphi}_S | \mathbf{B}_S) \pi(\mathbf{L}) d\mathbf{L} d\boldsymbol{\theta} d\boldsymbol{\beta} d\boldsymbol{\varphi}_S d\boldsymbol{\varphi}_C. \quad (12)$$

247 Samples from this distribution can be obtained by the following algorithm:

- 248 1. Sample  $N_1$  sets of catch locations  $\mathbf{L}$ .
- 249 2. Sample  $N_1$  sets of hyperparameters, regression coefficients and latent structures,  $\boldsymbol{\varphi}_S$ , from  
 250 the posterior distribution  $\pi(\boldsymbol{\theta}, \boldsymbol{\beta}, \boldsymbol{\varphi}_S | \mathbf{B}_S)$  using R-INLA.
- 251 3. Use the updating equations:

$$\begin{aligned} \mathbb{E}[\boldsymbol{\varphi}_C | \boldsymbol{\varphi}_S] &= \boldsymbol{\Sigma}_{CS} \boldsymbol{\Sigma}_{SS}^{-1} \boldsymbol{\varphi}_S \\ \text{Var}[\boldsymbol{\varphi}_C | \boldsymbol{\varphi}_S] &= \boldsymbol{\Sigma}_{CC} - \boldsymbol{\Sigma}_{CS} \boldsymbol{\Sigma}_{SS}^{-1} \boldsymbol{\Sigma}_{SC} \end{aligned} \quad (13)$$

252 to sample  $N_2$  realizations of  $\boldsymbol{\varphi}_C$  given  $\boldsymbol{\varphi}_S$  for each set of  $(\boldsymbol{\theta}, \boldsymbol{\beta}, \mathbf{L})$ .

- 253 4. For each sampled set of  $(\boldsymbol{\beta}, \boldsymbol{\varphi}_C, \boldsymbol{\theta})$  sample one value from  $\pi(\mathbf{B}_C | \boldsymbol{\beta}, \boldsymbol{\varphi}_C, \boldsymbol{\theta})$ .

254 The algorithm above samples  $N_1 N_2$  realizations of historical bycatch in the commercial fishery.

255 We selected  $N_1 = 100$  and  $N_2 = 50$  for the prediction of historical bycatch.

256 In Breivik et al. (2016) a prediction procedure implemented in R-INLA was used. Such a  
 257 prediction procedure could also have been used in this research, but then the full precision  
 258 matrix for the spatio-temporal Gaussian random field is required. We avoided working with this  
 259 large dense matrix by constructing a prediction procedure outside of R-INLA which only uses  
 260 sub-matrices of the full covariance matrix  $\boldsymbol{\Sigma}$ .

## 261 6 Prediction of historical bycatch

262 The object of this research is to predict the historical bycatch, and this result section is divided  
 263 into two subsections. The first subsection briefly shows the selected covariates and correlation  
 264 structures, and the second subsection shows the historical bycatch predictions of cod in the Bar-  
 265 ents Sea shrimp fishery. See appendix A.2 for details regarding the computational features.

### 266 6.1 Covariates and correlation

267 Table 4 lists covariates that were selected for prediction of bycatch. By inspecting the credibility  
 268 intervals, we found a clear effect of the 0-group. Furthermore, the inclusion of the 0-group halved  
 269 the variance of the year effect, leading to better predictive power, and is therefore included in the  
 270 model. As in Breivik et al. (2016), compared to using a single trawl, double trawl was shown to  
 271 increase bycatch while no effect was found for triple trawl. That triple trawl does not affect the  
 272 bycatch is intuitively surprising, and may be because only 3% of the survey data are collected  
 273 with use of triple trawl (see Table 2). Thereby may we not have enough observations to estimate  
 274 a possible triple trawl effect.

275 All three random terms in (5) were selected. This selection of random structure is the same as  
 276 in Breivik et al. (2016). See Table 4 for a summary of the estimated hyperparameters.

Covariates (eq. 5)			Hyperparameters		
Parameter	Mean	95% C.I.	Parameter	Mean	95% C.I.
Constant	-0.89	(-3.7,1.1)	$\varsigma$ (eq. 4)	2.09	(1.95,2.23)
depth (standardized)	-0.29	(-0.34,-0.25)	$a$ (eq. 6)	1.70	(1.53,1.88)
0-group	0.49	(0.21,0.76)	$\sigma_\alpha^2$ (eq. 8)	5.9	(2.2,14.8)
double trawl	0.43	(0.29,0.58)	$\kappa_\alpha$ (eq. 8)	0.0050	(0.0027,0.0078)
Shrimp catch (log scale)	0.36	(0.32,0.40)	$\sigma_v^2$	0.36	(0.11,0.87)
			$\sigma_\gamma^2$ (eq. 9)	1.9	(1.75,2.08)
			$\theta_t$ (eq. 9)	38 (mode)	unknown
			$\theta_s$ (eq. 9)	156 (mode)	unknown

Table 4: *Estimates and 95% credibility intervals of the significant regression coefficients and the hyperparameters.*

277 Figure 2 illustrates the spatial, seasonal and yearly effects for bycatch of cod. By comparing  
 278 the spatial contribution,  $\alpha(\mathbf{s})$  in equation (5), from Figure 2a with the juvenile cod migration  
 279 pattern in Jakobsen and Ozhigin (2011, page 227) we see a clear overlap. The seasonal effect,

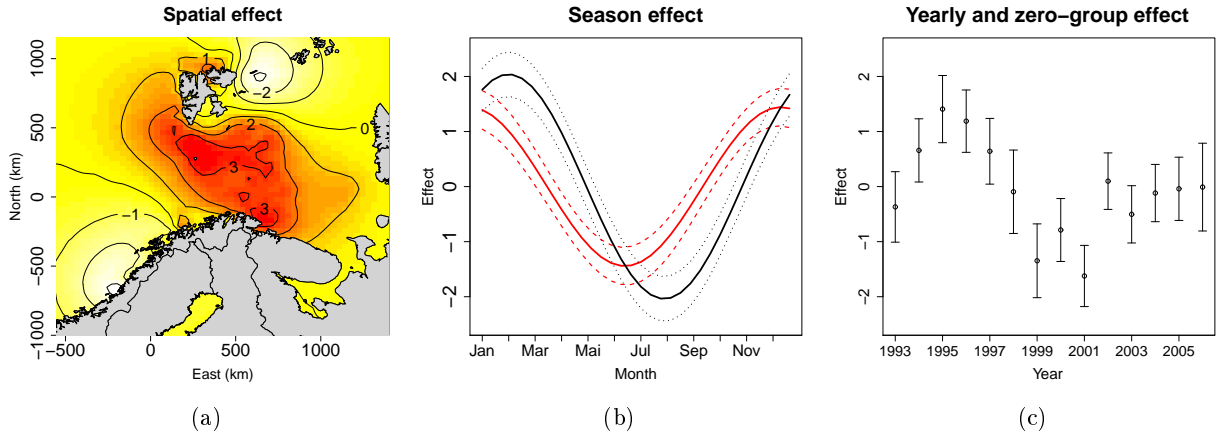


Figure 2: a) The spatial effect. b) The seasonal effect at 69 degrees north (red line) and 80 degrees north (black line) with 95% credibility intervals. c) The yearly effects added the zero-group effect with 95% credibility intervals, note that each interval illustrates the effect from 1st September in the denoted year to 31st August in the next year.

280 Figure 2b, is included with one harmonic in the Fourier series (7) and depends on latitude. Just  
 281 as in Breivik et al. (2016), the seasonal effect increases later in autumn in the north compared  
 282 to in the south, see Figure 2b.

## 283 6.2 Prediction

284 This subsection presents the predicted number of juvenile cod killed as bycatch each year in  
 285 the Barents Sea shrimp fishery. Our predictions are reported with posterior means and 90%  
 286 prediction intervals. The predicted yearly historical bycatch (with uncertainty) is shown in  
 287 Figure 3. The predicted yearly historical bycatch with quarterly predictions are further given  
 288 in Table 5. There seems to be variation between years, which is reasonable since the fishing  
 289 intensity and the cod year class strength changes from year to year.

290 In addition, Figure 3 includes historical bycatch estimates with the ratio method (red crosses)  
 291 and effort method (green triangles). We see that our method is often in agreement with the  
 292 ratio and effort methods, but clearly differed from the ratio method in year 1998 and 2004. A  
 293 main reason why they differ is because of the sensitivity of the ratio method to small shrimp  
 294 catches. In the fourth quarter of year 2004 there were five observations in the survey data  
 295 which lead to a bycatch ratio of 38.9 in a specific area north in the Barents Sea. In this area  
 296 the commercial fishery was 128 times more efficient than the MSS to catch shrimp per hour of

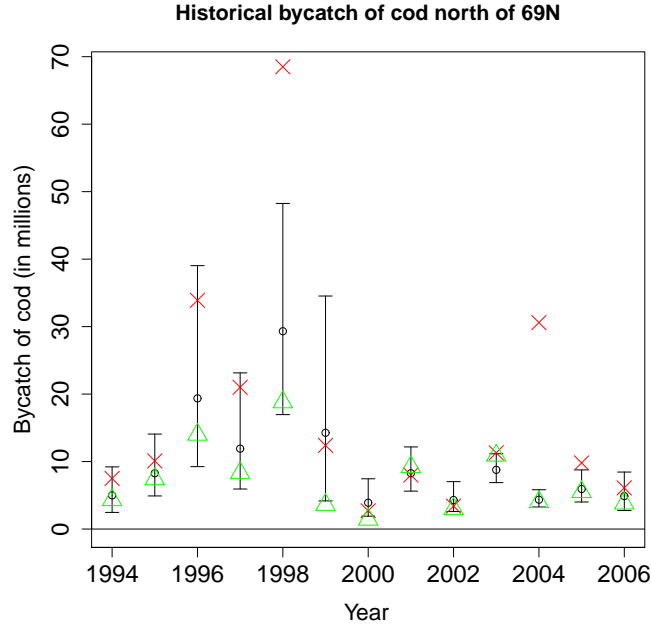


Figure 3: *Posterior means of yearly historical bycatch with 90% prediction intervals. The red crosses are the ratio estimates (1) and the green triangles are the effort estimates (2).*

297 trawl, which implies that the ratio estimate was not representative for the commercial fishery.  
 298 Removing these five observations resulted in a ratio method estimate of 3.9 million instead of  
 299 30.6 million cod in year 2004, which is much more in agreement with our model-based approach.  
 300 The difference in year 1998 can be explained likewise, and is omitted for brevity. The effort  
 301 method (2) is not sensitive to small shrimp catches since it neglects the target catch, but is  
 302 however sensitive to short trawl hauls.

303 Note that depth is included as a covariate in the prediction procedure, while not given in the  
 304 fishery data (see Table 1). The depth at the commercial catch location is in this research  
 305 extrapolated to be the same as the depth at the closest surveillance observation in space for  
 306 prediction. The survey data are concentrated where commercial shrimp trawling occurs, and we  
 307 therefore assume this approximation is sufficient.

## 308 7 Validation

309 In this section we validate the models ability to produce reliable bycatch predictions with uncer-  
 310 tainty. This validation section is divided into three subsections. The first subsection validates  
 311 predictions and uncertainty estimates of aggregated bycatch. The second subsection validates



Year	Total	1st quarter	2nd quarter	3rd quarter	4th quarter	Shrimp catch
1994	5.0 (2.5,9.2)	2.5 (0.8,5.7)	0.7 (0.3,1.3)	0.9 (0.2,2.5)	0.9 (0.3,1.9)	18900 tons
1995	8.3 (4.9,14.1)	2.9 (1.3,6.1)	2.5 (1.6,3.7)	1.7 (0.4,4.8)	1.2 (0.2,3.4)	15600 tons
1996	19.4 (9.2,39.0)	6.4 (1.0,19.3)	8.0 (3.5,17.1)	4.2 (2.2,7.6)	0.7 (0.2,1.7)	20500 tons
1997	11.9 (5.9,23.1)	2.6 (0.7,6.6)	4.8 (2.2,10.4)	3.5 (1.0,9.1)	1.0 (0.3,2.6)	25600 tons
1998	29.3 (17.0,48.3)	17.7 (8.4,32.9)	7.6 (4.0,13.0)	2.6 (0.6,6.8)	1.5 (0.3,4.0)	41200 tons
1999	14.3 (4.2,34.5)	7.5 (1.3,21.7)	4.4 (1.0,12.0)	2.0 (0.4,5.4)	0.3 (0.1,0.5)	48400 tons
2000	3.9 (1.9,7.4)	1.9 (0.5,5.0)	0.6 (0.3,1.0)	0.8 (0.3,2.0)	0.5 (0.2,1.3)	52000 tons
2001	8.3 (5.6,12.2)	2.8 (1.6,4.8)	2.7 (1.5,4.7)	1.2 (0.4,2.8)	1.5 (0.9,2.5)	42200 tons
2002	4.3 (2.6,7.0)	2.3 (0.8,4.8)	1.1 (0.7,1.7)	0.2 (0.1,0.4)	0.7 (0.4,1.2)	49500 tons
2003	8.8 (6.9,11.2)	0.7 (0.3,1.2)	5.0 (3.6,6.9)	2.8 (2.0,4.0)	0.3 (0.1,0.7)	33200 tons
2004	4.4 (3.3,5.8)	1.4 (0.8,2.2)	1.8 (1.2,2.5)	0.7 (0.4,1.1)	0.5 (0.3,0.9)	35000 tons
2005	5.9 (4.0,8.8)	1.4 (0.8,2.5)	2.2 (1.3,3.6)	1.8 (0.9,3.2)	0.5 (0.2,1.2)	34000 tons
2006	4.9 (2.7,8.4)	1.5 (0.4,4.0)	2.5 (1.3,4.4)	0.3 (0.2,0.6)	0.5 (0.1,1.5)	27900 tons

Table 5: *Yearly and quarterly historical bycatch predictions of cod with 90% prediction intervals (in millions), and yearly aggregated Norwegian commercial shrimp catch.*

312 model assumptions. The third section investigates prediction bias and power using a simulation  
313 study. Due to the computational cost of integrating out the uncertainty in the hyperparameters,  
314 validation is performed with empirical Bayes, i.e. using posterior mode of hyperparameters. We  
315 have observed that the bycatch predictions are typically little affected by using the posterior  
316 mode of the hyperparameters, which indicates that this procedure does not strongly influence  
317 the validation.

## 318 7.1 Validation of predictions

319 This subsection validates the predictions, and shows that the model is able to give realistic  
320 predictions and uncertainty measures. The fishery and survey data are typically clustered in  
321 space and time. Therefore, to make the validation representative for the prediction purpose, the  
322 survey data are divided into clustered training and test sets. The clustering is accomplished by  
323 first dividing the survey data into *fishing trips*. A fishing trip is here defined as the largest set  
324 of observations taken by one distinct boat such that every time gap between two observations  
325 next to each other in chronological order is less than 3 days. The clustered test sets are then  
326 constructed with the same reasoning as in Hastie et al. (2009, page 241) by uniformly dividing  
327 the fishing trips into ten groups with equally many fishing trips within each group. Each group

328 is then used as a test set and the others as the training set. This procedure is repeated 100  
 329 times leading to in total 1000 test and training sets. Note that we only use the survey data for  
 330 validation of predictions since we know the true observed bycatch in the survey data, and can  
 331 thereby compare the predictions with the truth.

332 Figure 4 shows predicted aggregated bycatch in the test sets versus the true observation with  
 333 Bayesian p-values (Gelman et al., 2003, page 162). We see from Figure 4a that our model has  
 334 predictive power, and by inspection of the Bayesian p-values in Figure 4b we observe that the  
 335 model is able to give reasonable uncertainty estimates (since the p-values are roughly uniformly  
 336 distributed). The relatively few small Bayesian p-values in Figure 4b indicates that the upper  
 337 bound of the prediction intervals of historical bycatch in Figure 3 and in Table 5 might be slightly  
 338 overestimated. Figure 7 illustrates the Bayesian p-values if we neglect parts of the random  
 339 effects in the model (5), and we observe that the random effects are crucial for estimating the  
 340 uncertainty, properly.

341 Coverage of bycatch predictions in the test sets in three common prediction interval levels are  
 342 given in Table 6. Just as in Figure 4b, we observe that our model typically overestimate the  
 343 upper bound of the prediction intervals.

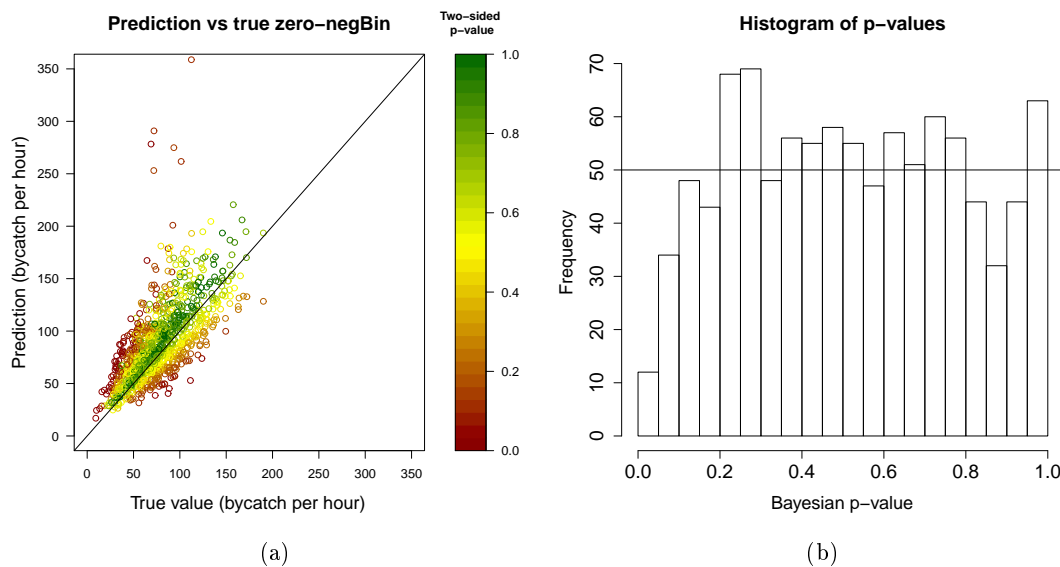


Figure 4: *a) Plot of predicted bycatch versus observed bycatch per hour trawl in the test sets, with color code illustrating the two sided p-values. b) Histogram of the Bayesian p-values. The horizontal line show the expected frequency of p-values if the model was correct.*

344 The accuracy of the prediction procedures is investigated with the mean absolute *relative error*

P.I. level	Inside P.I.	Under P.I.	over P.I.
90%	92.4%	6.2%	1.3%
95%	95.6%	4.0%	0.4%
99%	98.4%	1.2%	0.2%

Table 6: Coverage of our model in three common prediction interval levels.

345 of aggregated bycatch in the test sets. The relative error is defined as

$$\text{relative error} = \frac{\text{prediction} - \text{true value}}{\text{true value}}. \quad (14)$$

346 With the ratio method, effort method and our model based approach the mean absolute relative  
347 error is equal to 0.51, 0.34 and 0.32 respectively. This indicates that our prediction procedure  
348 is more accurate than the ratio method which is currently in use for providing official historical  
349 bycatch estimates in the Barents Sea shrimp fishery.

350 The two range parameters in the spatio-temporal interaction (9) are estimated with all the  
351 survey data (that is both the training and test set) when predicting bycatch in the test sets.  
352 This was done due to the computational cost of estimating these parameters. We have observed  
353 that the posterior mode of the range parameters in the spatio-temporal interaction is approxi-  
354 mately unchanged when estimated with several different training sets, which indicates that this  
355 procedure does not influence the validation of prediction.

## 356 7.2 Validation of model assumptions

357 Model assumptions are investigated using Pearson type residuals (McCullagh and Nelder, 1989,  
358 page 37) as recommended in Zuur and Ieno (2016). The residuals are calculated by sequen-  
359 tially leaving out every tenth surveillance observations and predicting them. Plots of Pearson  
360 residuals versus time and space coordinates and versus explanatory variables are investigated  
361 for correlation structures and for evidence of non-linearity in (5), and no clear violations are  
362 observed. All these plots are given in the online supplementary information. We also include  
363 Pearson residuals plotted against the order of each continuous variable, these are included to  
364 make clustered Pearson residuals easier to validate visually. As an example, Figure 5 shows

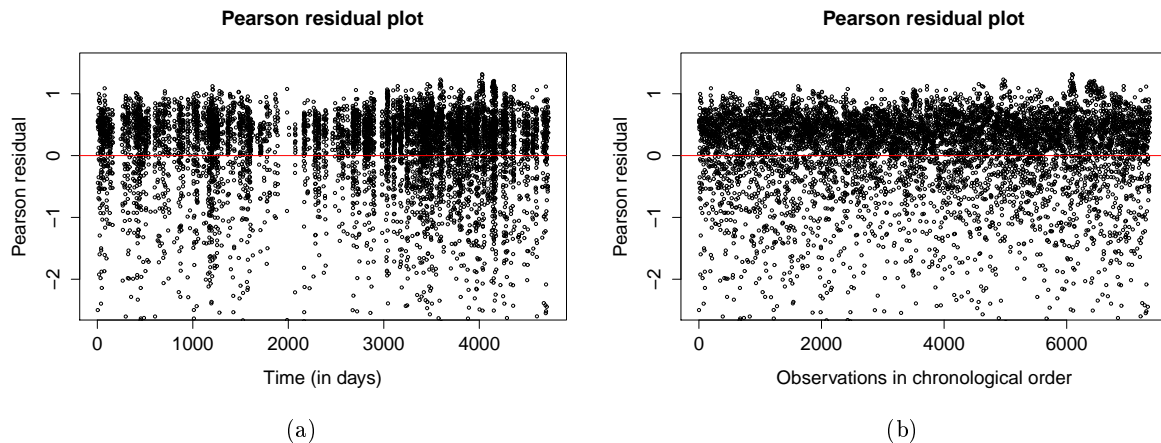


Figure 5: a) *Pearson residuals versus time.* b) *Pearson residuals in chronological order.*

365 Pearson residuals plotted against time. Variogram and autocorrelation plots are included in the  
 366 online supplementary information, and give no indication of violations.

### 367 7.3 Validation through a simulation study

368 In this subsection we investigate the bias of historical bycatch predictions when assuming our  
 369 model describes the underlying stochastic structure of the bycatch observations. The ratio and  
 370 effort method are observed to be typically biased, while no such structure is observed for the  
 371 model-based procedure. The validation is conducted by first simulating bycatch conditioned on  
 372 the observed shrimp catch (only 10% of the fishery data from each year, chosen at random, is  
 373 used due to computation time). See appendix A.3 for a description of the joint simulation of  
 374  $\mathbf{B}_C$  and  $\mathbf{B}_S$ . The bias is then investigated through the distribution of the relative error (14) of  
 375 the aggregated simulated commercial bycatch.

376 A boxplot summary of 100 simulated relative errors of aggregated yearly bycatch in the com-  
 377 mercial fishery is shown in Figure 6. We see that there is a tendency to overestimate bycatch  
 378 when using the ratio method (Figure 6a), and a tendency to underestimate when using the  
 379 effort method (Figure 6b). This bias can be explained by that the commercial fishery focuses on  
 380 areas with high density of shrimps, while survey data are relatively random located where shrimp  
 381 trawling occurs. Our research indicates (Table 4) that a doubling of shrimp catch (given un-  
 382 changed trawling effort) imply a bycatch increase of approximately 28%, while the ratio (1) and  
 383 effort (2) methods on the other hand assumes 100% and 0% increase respectively. Given that

384 the commercial fishery catches shrimps more effectively than the MSS, this indicates that the  
 385 ratio method typically overestimates while the effort method typically underestimates historical  
 386 bycatch.

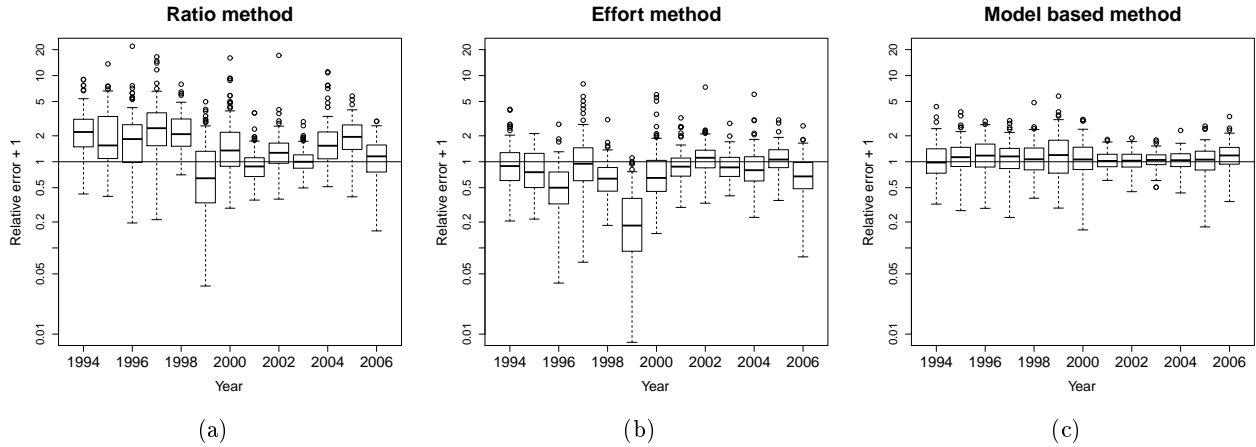


Figure 6: *Illustration of relative error with the ratio method (a), with the effort method (b) and with our model based approach (c). Note that the y-axis is on logarithmic scale.*

387 Figure 6c illustrates the relative error when using our model based approach. Given that our  
 388 model represents the true underlying stochastic structure, we observe that it gives reasonable  
 389 unbiased predictions and thereby has predictive power.

390 “When simulating data from the model, the simulated data should be comparable to the original  
 391 data. If not, the model needs improvement” (Zuur and Ieno, 2016). By investigating the  
 392 simulations with the true observed bycatch, with respect to number of zeros, maximum value,  
 393 total bycatch, median bycatch and visual inspection, we observed that they are comparable (see  
 394 online supplementary information for details).

## 395 8 Discussion

396 The object of this research has been to predict historical bycatch in commercial fishery by using a  
 397 Bayesian spatio-temporal latent Gaussian model. This discussion is divided in three parts. First  
 398 we discuss the importance of random effects in our model. Secondly we discuss the observation  
 399 model used. Thirdly we compare the historical bycatch with abundance estimates of cod.

## 400 **8.1 The importance of random effects**

401 Predictions of bycatch using model-based procedures has been conducted earlier. Murray (2005)  
402 used a generalized additive model to predict the total bycatch of loggerhead turtles in the  
403 Atlantic Sea scallop dredge fishery without random effects. Pennino et al. (2014) investigated  
404 a spatio-temporal model for bycatch without the spatio-temporal interaction. Figure 7 shows  
405 the estimated p-value of aggregated bycatch in the test sets if we use no random effect or  
406 a spatio-temporal structure without spatio-temporal interaction respectively. By comparing  
407 Figure 7 with Figure 4b we see that the model including all selected random effects much better  
408 estimates the uncertainty since the Bayesian p-values are more uniformly distributed.

409 Cosandey-Godin et al. (2014); Ward et al. (2015) investigated spatio-temporal models for by-  
410 catch with a separable spatio-temporal interaction function that discretizes time and uses an  
411 autocorrelated structure of order one in time and a Matern correlation structure in space. Such  
412 a discretized spatio-temporal structure was also considered with the survey data in Breivik  
413 et al. (2016), but the continuous correlation function (9) was favored and therefore used in this  
414 research. A problem encountered with the spatio-temporal correlation function in Cosandey-  
415 Godin et al. (2014) is that our data are unstructured and a coarse grid in both space and time  
416 is needed for the model to be computationally feasible due to the large imposed grid structure  
417 in space and time (Cameletti et al., 2013). We have predicted the historical bycatch in several  
418 years with use of the spatio-temporal interaction function in Cosandey-Godin et al. (2014) (with  
419 time discretized in 30 days, and with spatial locations more than 80 km from each other in  
420 the spatial grid) and the predictions were similar to ours most of the years (not shown). Some  
421 years however were predicted different, but by using finer temporal discretization (20 days), the  
422 predictions were more similar. This is not surprising since a relatively fine temporal and spatial  
423 discretization results in that the spatio-temporal interaction structure in Cosandey-Godin et al.  
424 (2014) is approximately similar to the one used in this research (Breivik et al., 2016).

## 425 **8.2 Survey data compared with fishery data**

426 This research utilizes two data sources, survey data and fishery data, and it is assumed that the  
427 survey data are representative for the fishery data for predicting bycatch *given shrimp catch*.  
428 In fisheries research it is commonly assumed that expected catch is expressed as a product of

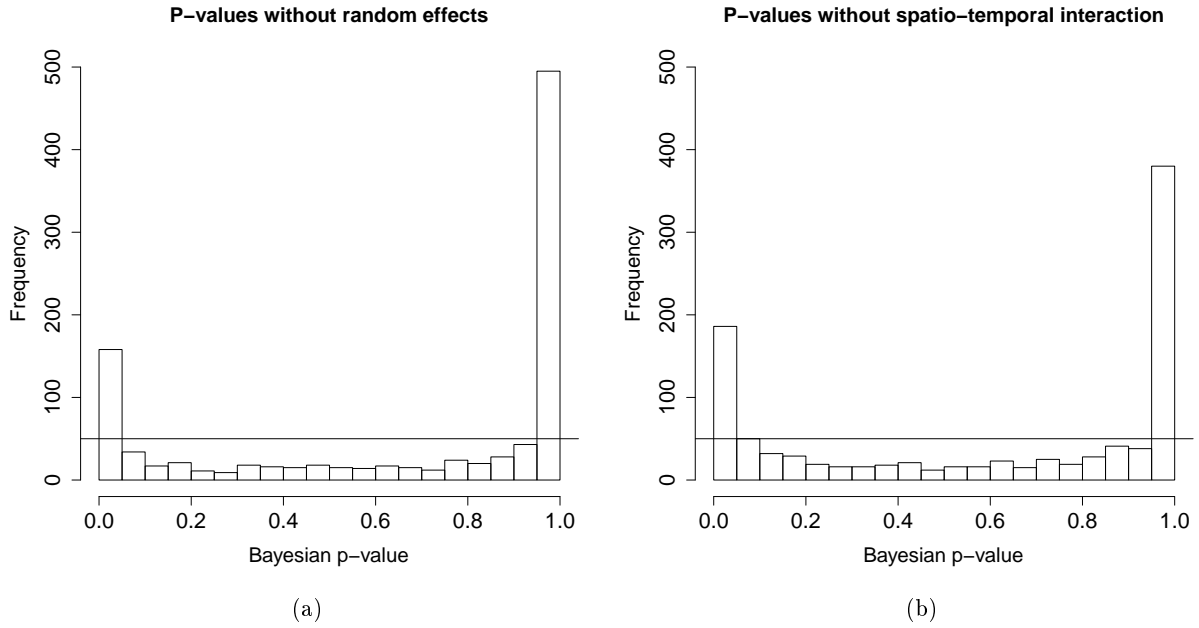


Figure 7: Bayesian  $p$ -values of hourly bycatch in the test sets without using random effects (a) and with spatial and temporal random effects but without the spatio-temporal interaction (b). The horizontal line show the expected frequency of  $p$ -values if the model was correct.

429 the catchability and the local density of the species (Thorson et al., 2016). The survey data  
 430 are collected using the same type of equipment as used in the commercial fishery. Thereby, we  
 431 argue that the assumption of representative catchability is reasonable. The density of bycatch is  
 432 indifferent of the purpose of the trawl. However, some of the survey observations are taken due  
 433 to expected high bycatch ratios of a commercial fish species or of undersized shrimps, e.g. due  
 434 to seasonal effects or information received by the fishery (MSS, pers. comm.). The commercial  
 435 fishery may also behave differently when an observer is on board, e.g. to avoid high bycatch  
 436 ratios for saving time and fuel needed to leave a closed area. The presence of observations taken  
 437 due to information not included in the analysis (e.g. the fisheries knowledge about the spatio-  
 438 temporal interaction effect for cod) may introduce a bias in the predictions. This possible bias is  
 439 assumed to be small, and is neglected in our analysis. Note that the MSS regulates with respect  
 440 to several other fish species, as described in section 1. These species have different juvenile  
 441 migration patterns compared to cod (Jakobsen and Ozhigin, 2011), which is an argument for  
 442 why such a possible bias introduced should be small. We want to emphasize that the procedure  
 443 used in this research should be generalized to other fisheries with caution if there are reasons to  
 444 question the assumption of representative survey data.

445 The exact spatial locations of the fishery data are not given, which differs from the survey data.

446 To accommodate for the uncertainty in location, the commercial catch locations are sampled  
447 uniformly within the areas reported (see green rectangles in Figure 1). It is reasonable that the  
448 catch locations are clustered in both time and space, which typically increases the uncertainty  
449 of the predictions through the spatio-temporal interaction. However, we assume that this effect  
450 is small and neglect it in our analysis. Note further that the commercial catches are reported as  
451 daily catches, meaning that two separate catches are treated as *one* if they are caught the same  
452 day and in the same area. This differs from the survey data, where each catch is distinctly given.  
453 That the commercial bycatch is modeled with daily catches may introduce an overestimation of  
454 the uncertainty.

### 455 **8.3 Observation models**

456 Breivik et al. (2016) models bycatch with use of a log-Gaussian observation model. However,  
457 O’hara and Kotze (2010); Warton et al. (2016) make a strong case that counting data should  
458 be modeled with a counting distribution rather than a log-Gaussian. After a comment from a  
459 reviewer, a zero-inflated negative binomial observation model was therefore investigated in this  
460 research. By comparing the predictions of aggregated bycatch in the test sets in section 7.1, the  
461 zero-inflated negative binomial model was favored due to a clear observed underestimation by  
462 the log-Gaussian model. The removal of this underestimation is a main reason for modifying  
463 the model in Breivik et al. (2016) to a zero-altered negative binomial model. Since we use the  
464 user-friendly R-package R-INLA, such a change of data distribution is easily achieved by only  
465 changing a few lines in the R-code. However, the non-Gaussian data distribution results in a  
466 more complex and time consuming inference of the latent structure, especially when utilizing  
467 the uncertainty in the hyperparameters (Rue et al., 2009).

### 468 **8.4 The impact of bycatch on the cod population**

469 This subsection compares estimated abundance of one year old cod with the predicted historical  
470 bycatch. Figure 8 shows the total historical bycatch of cod in each year as a percentage of  
471 the estimated aggregated abundance of one year old cod obtained from (ICES, 2015). This  
472 figure might give a rough indication on how much bycatch is caught compared with aggregated  
473 abundance estimated in the beginning of the year. Note, however, that the uncertainty in the



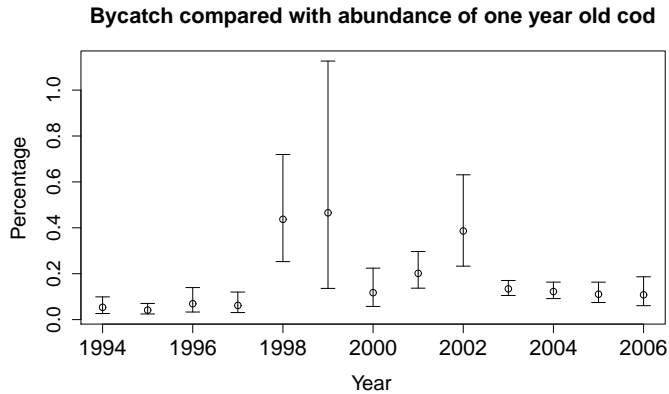


Figure 8: *Historical bycatch as percentage of the estimated aggregated abundance of one year old cod (ICES, 2015) in the Barents Sea. The intervals represent 90% prediction intervals when neglecting the uncertainty in the abundance estimates.*

474 abundance estimates are not given in ICES (2015), and therefore should the prediction intervals  
 475 given in Figure 8 be wider (these are only based on uncertainty in the bycatch predictions).  
 476 Note further that there is a regulation regime in the Barents Sea which closes areas when high  
 477 bycatch ratios are observed, and without the regulation regime the historical predictions could  
 478 have been larger. The relative low total bycatch may hence illustrate the success of the current  
 479 regulation regime.

## 480 9 Conclusions and further work

481 We conclude that the model-based procedure produces reliable predictions (including uncertainty  
 482 measures) of historical cod bycatch in the Barents Sea shrimp fishery, see section 7.1. We further  
 483 make a strong case that the Bayesian spatio-temporal model based method outperforms both  
 484 the ratio and effort methods for prediction of historical bycatch. This argument is based on the  
 485 following observations elaborated in the article:

- 486 • The ratio and effort methods are sensitive to small shrimp catches and short trawl hauls  
 487 respectively, see section 6.2.
- 488 • The model based method produces reliable predictions with uncertainty estimates, see  
 489 section 7.1.
- 490 • The shrimp catch is positively correlated with bycatch (Table 4), indicating that both the

491 ratio and effort methods are biased, see section 7.3.

492 Further work is desirable on prediction of historical bycatch for other species and in other  
493 fisheries to investigate the generality of the model based approach. We strongly believe similar  
494 spatio-temporal models are useful for bycatch predictions of other species and in other fisheries.  
495 The R-code used for predicting bycatch is available upon request.

## 496 Acknowledgments

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498 also very thankful for constructive discussion with several employees at the Norwegian Institute  
499 of Marine Research and the Norwegian Directorate of Fisheries Monitoring and Surveillance  
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501 reviewers who gave constructive comments and suggestions which improved the article.

## 502 Appendix

### 503 A.1 Priors

504 The priors for the hyperparameters used in this research are given in Table A.1. These are  
505 constructed to be relatively non-informative. The gamma distribution used has the parametriza-  
506 tion:

$$\pi(x|\alpha,\beta) = \frac{\beta^\alpha}{\Gamma(\alpha)} x^{\alpha-1} \exp(-\beta x). \quad (\text{A.1})$$

507 R-INLA by default uses an improper prior for the intercept regression coefficient and a  $N(0,1000)$   
508 distribution for the other regression coefficients.

Parameter	Prior	Parameter	Prior
$\log(\sigma_\alpha^2)$	N(0,10)	$1/\sigma_\nu^2$	gamma(1,0.00005)
$\log(\kappa)$	N(0,10)	$1/\sigma_\gamma^2$	gamma(1,0.00005)
$\log(\varsigma)$	N(1,1)	$\theta_t$ and $\theta_s$	$\propto 1$
$\log(a)$	N(2,1)		

Table A.1: *Prior distributions.*

## 509 A.2 Computational features

510 The first step of our historical bycatch prediction procedure is to estimate the parameters in  
511 the model given the survey data. This took approximately 1.4 hours on an Intel Core i5-2500  
512 CPU 3.30GHz  $\times$  4 processor (with good starting values of the Newton method used to find  
513 posterior mode of the hyperparameters within R-INLA and after the posterior mode of the range  
514 parameters in the spatio-temporal interaction was found). The second part of the predicting  
515 procedure of historical bycatch is done on a cluster of computers. Notice the parallel structure  
516 caused by the independent simulation of catches. We used 20 cores each with 32 gigabyte  
517 memory and 2.20GHz. This second part took 1.5 hour to 5 hours for each year, depending on  
518 the number of daily catches.

## 519 A.3 Joint simulation of $\mathbf{B}_C$ and $\mathbf{B}_S$

520 This section elaborates the joint simulation procedure for commercial bycatch and bycatch in  
521 survey data. The simulation is done with the following algorithm:

- 522 1. Find the posterior mode of the hyperparameters,  $\hat{\boldsymbol{\theta}}$ , given  $\mathbf{B}_S$ .
- 523 2. Sample  $\boldsymbol{\beta}^*$  and  $\boldsymbol{\alpha}^* = \{\alpha^*(\mathbf{s})\}$  from  $\pi(\boldsymbol{\beta}, \boldsymbol{\alpha} | \mathbf{B}_S, \hat{\boldsymbol{\theta}})$ .
- 524 3. Sample  $\mathbf{B}_C^*$  and  $\mathbf{B}_S^*$  from  $\pi(\mathbf{B}_C, \mathbf{B}_S | \hat{\boldsymbol{\theta}}, \boldsymbol{\beta}^*, \boldsymbol{\alpha}^*)$ .

525 Notice that we use the full posterior distribution of the regression coefficients and the spatial  
526 effect while we only use the posterior mode of the hyperparameters.

## 527 **References**

- 528 Ajiad, A., Aglen, A., Nedreaas, K., and Kvamme, C. 2007. NAFO/ICES Pandalus Assessment  
529 Group Meeting.
- 530 Amandè, M. J., Ariz, J., Chassot, E., De Molina, A. D., Gaertner, D., Murua, H., Pianet, R.,  
531 Ruiz, J., and Chavance, P. 2010. Bycatch of the European purse seine tuna fishery in the  
532 Atlantic Ocean for the 2003–2007 period. *Aquatic Living Resources*, 23(4):353–362.
- 533 Blangiardo, M. and Cameletti, M. 2015. *Spatial and Spatio-temporal Bayesian Models with*  
534 *R-INLA*. John Wiley & Sons.
- 535 Breivik, O. N., Storvik, G., and Nedreaas, K. 2016. Latent Gaussian models to decide on spatial  
536 closures for bycatch management in the Barents Sea shrimp fishery. *Canadian Journal of*  
537 *Fisheries and Aquatic Sciences*, 73(8):1271–1280.
- 538 Cameletti, M., Lindgren, F., Simpson, D., and Rue, H. 2013. Spatio-temporal modeling of  
539 particulate matter concentration through the SPDE approach. *AStA Advances in Statistical*  
540 *Analysis*, 97(2):109–131.
- 541 Cosandey-Godin, A., Krainski, E. T., Worm, B., and Flemming, J. M. 2014. Applying Bayesian  
542 spatiotemporal models to fisheries bycatch in the Canadian Arctic. *Canadian Journal of*  
543 *Fisheries and Aquatic Sciences*, 72(2):186–197.
- 544 Davies, R., Cripps, S., Nickson, A., and Porter, G. 2009. Defining and estimating global marine  
545 fisheries bycatch. *Marine Policy*, 33(4):661–672.
- 546 Fiskeridirektoratet 2005. Forskrift om utøvelse av fisket i sjøen (in Norwegian).  
547 <https://lovdata.no/dokument/SF/forskrift/2004-12-22-1878> (last accessed September 18,  
548 2016).
- 549 Gelfand, A. E. 1996. Model determination using sampling-based methods, pages 145–161. Lon-  
550 don: Chapman and Hall.
- 551 Gelman, A., Carlin, J. B., Stern, H. S., and Rubin, D. B. 2003. *Bayesian Data Analysis*.  
552 Chapman and Hall/CRC, 2 edition.
- 553 Hall, M. A. 1996. On bycatches. *Reviews in Fish Biology and Fisheries*, 6(3):319–352.

554 Hastie, T., Tibshirani, R., Friedman, J., Hastie, T., Friedman, J., and Tibshirani, R. 2009. The  
555 elements of statistical learning, volume 2. Springer.

556 Hylan, A. and Jacobsen, J. 1987. Estimation of cod taken as by-catch in the norwegian fishery  
557 for shrimp north of 69 N. ICES CM.

558 ICES 1994. Report of the Arctic Fisheries Working group, Copenhagen, 24 August - 2 September  
559 1993.

560 ICES 2015. Report of the Arctic Fisheries Working Group (AFWG), 23-29 April 2015 Hamburg,  
561 Germany.

562 Jakobsen, T. and Ozhigin, V. K. 2011. The Barents Sea-ecosystem, resources, management.  
563 Half a century of Russian-Norwegian cooperation. Tapir Akademisk Forlag.

564 Lay, D. C. 2006. Linear Algebra and Its Applications, Third Edition. Person.

565 Little, A. S., Needle, C. L., Hilborn, R., Holland, D. S., and Marshall, C. T. 2015. Real-time  
566 spatial management approaches to reduce bycatch and discards: experiences from Europe and  
567 the United States. *Fish and Fisheries*, 16(4):576–602.

568 Martins, T. G., Simpson, D., Lindgren, F., and Rue, H. 2013. Bayesian computing with INLA:  
569 New features. *Computational Statistics & Data Analysis*, 67:68–83.

570 McCullagh, P. and Nelder, J. A. 1989. Generalized linear models, volume 37. CRC press.

571 Murray, K. 2005. Total bycatch estimate of loggerhead turtles (*Caretta caretta*) in the 2004  
572 Atlantic sea scallop (*Placopecten magellanicus*) dredge fishery. US Dep Commer, Northeast  
573 Fish Sci Cent Ref Doc, pages 05–12.

574 O’hara, R. B. and Kotze, D. J. 2010. Do not log-transform count data. *Methods in Ecology and*  
575 *Evolution*, 1(2):118–122.

576 Pennino, M. G., Muñoz, F., Conesa, D., López-Quílez, A., and Bellido, J. M. 2014. Bayesian  
577 spatio-temporal discard model in a demersal trawl fishery. *Journal of Sea Research*, 90:44–53.

578 Rue, H., Martino, S., and Chopin, N. 2009. Approximate Bayesian inference for latent Gaussian  
579 models by using integrated nested Laplace approximations. *Journal of the Royal Statistical*  
580 *Society: Series B (Statistical Methodology)*, 71(2):319–392.

- 581 Scheaffer, R., Mendenhall III, W., and Ott, R. L. 1996. Elementary survey sampling, Fifth  
582 Edition. Duxbury Press.
- 583 Thorson, J. T., Fonner, R., Haltuch, M. A., Ono, K., and Winker, H. 2016. Accounting for  
584 spatio-temporal variation and fisher targeting when estimating abundance from multispecies  
585 fishery data. *Canadian Journal of Fisheries and Aquatic Sciences*, (in press).
- 586 Vinther, M. 1999. Bycatches of Harbour Porpoises (*Phocoena phocoena*, L.) in Danish set-net  
587 fisheries. *Journal of Cetacean Research and Management*, 1(2):123–135.
- 588 Walmsley, S. A., Leslie, R. W., and Sauer, W. H. 2007. Bycatch and discarding in the South  
589 African demersal trawl fishery. *Fisheries Research*, 86(1):15–30.
- 590 Ward, E. J., Jannot, J. E., Lee, Y.-W., Ono, K., Shelton, A. O., and Thorson, J. T. 2015. Using  
591 spatiotemporal species distribution models to identify temporally evolving hotspots of species  
592 co-occurrence. *Ecological Applications*, 25(8):2198–2209.
- 593 Warton, D. I., Lyons, M., Stoklosa, J., and Ives, A. R. 2016. Three points to consider when  
594 choosing a LM or GLM test for count data. *Methods in Ecology and Evolution*, (7):882–890.
- 595 Ye, Y. 2002. Bias in estimating bycatch-to-shrimp ratios. *Aquatic Living Resources*, 15(03):149–  
596 154.
- 597 Ye, Y., Alsaffar, A., and Mohammed, H. 2000. Bycatch and discards of the Kuwait shrimp  
598 fishery. *Fisheries Research*, 45(1):9–19.
- 599 Zuur, A. and Ieno, E. 2016. A protocol for conducting and presenting results of regression-type  
600 analyses. *Methods in Ecology and Evolution*, 7(6):636–645.
- 601 Zuur, A. F. 2009. *Mixed effects models and extensions in ecology with R*. Springer.